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Theoretical predictions of arsenic and selenium species under atmospheric conditions

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Abstract

Thermochemical properties of arsenic and selenium species thought to be released into the atmosphere during the coal combustion were examined using ab initio methods. At various levels of theory, calculated geometries and vibrational frequencies of the species were compared with experimental data, where available. Through a comparison of equilibrium constants for a series of gaseous arsenic and selenium oxidation reactions involving OH and HO₂, five thermodynamically favored reactions were found. In addition, it was determined that all favored reactions were more likely to go to completion under tropospheric, rather than stratospheric, conditions.

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1. Introduction

Globally, coal combustion provides a major source of energy with more than five billion short tons of coal burned each year [\(Katzer, 2007\)](#page-7-0). In the United States, approximately 52% of electricity comes from coal, an amount expected to increase in the coming decades ([Energy Information Adminis](#page-7-0)[tration \(EIA\), 2005;](#page-7-0) [Katzer, 2007](#page-7-0)). As the use of coal for power increases, attention turns to the potentially harmful effects associated with its combustion. Along with the well-known problem of CO2 emissions, various trace elements are emitted. When coal combusts, both arsenic and

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selenium are volatilized and escape the smokestacks either as a gas or enriched in the fly ash ([Senior](#page-7-0) [et al., 2000](#page-7-0)). Both trace elements naturally occur in the earth's crust; however, at high concentrations they can have adverse effects on the environment and human health. By understanding the speciation and lifetime of arsenic and selenium in the atmosphere, technologies can be developed to better deal with these potentially harmful elements.

Arsenic ranks 20th in trace elements found in the earth's crust ([National Research Council \(NRC\),](#page-7-0) [1997\)](#page-7-0). Natural cycles, such as weathering and volcanic activity, emit arsenic into the atmosphere [\(Cullen and Reimer, 1989](#page-7-0)); however, studies have traced nearly 75% of arsenic emissions to anthropogenic sources ([Zeng et al., 2001\)](#page-8-0). Arsenic content in coal averages 2 ppm ([Commission on Life](#page-7-0) [Sciences \(CLS\), 1977](#page-7-0)), making coal combustion

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a major source of anthropogenic arsenic release [\(Han et al., 2003](#page-7-0)). When arsenic species enter drinking water sources humans may suffer arsenic toxicity, symptoms of which include problems of the gastrointestinal tract, circulatory system, immune system, and renal system. Skin lesions and cancers, particularly Bowen's disease, may also result [\(Duker](#page-7-0) [et al., 2005](#page-7-0)). To protect drinking water sources it is important to look at the lifetime of arsenic species in the atmosphere, i.e., whether they enter the stratosphere, are flushed out with rainwater, or adsorb onto dust particles and settle into soil. Arsenic exists in two oxidation states, the trivalent As(III) and pentavalent As(V). As(III) molecules, such as $AsCl₃$ and AsH3, tend to be water-soluble [\(Hudson-](#page-7-0)[Edwards et al., 2005](#page-7-0); [Efermov et al., 2002\)](#page-7-0). The species $As₂$, $As₄$, and elemental As are also found to be at least partially water-soluble [\(Young, 1979;](#page-8-0) [Harrington et al., 1997\)](#page-7-0). If these molecules are stable in the atmosphere, they will likely end up in soil or fresh water. Arsenic (V) molecules tend to be prone to adsorption ([Frentiu et al., 2007](#page-7-0)) and will likely make their way into soil by adsorbing onto dust or soot particulates.

Selenium levels in coal average 1 ppm [\(Shah et al.,](#page-7-0) [2007](#page-7-0)), making coal combustion the main source of anthropogenic release of selenium [\(ATSDR, 2003](#page-8-0)). As with arsenic, selenium uptake by humans can be toxic, but, unlike arsenic, selenium plays an essential role as an antioxidant enzyme [\(Tinggi, 2003\)](#page-8-0). The lower recommended daily intake of selenium is 55μ g day⁻¹ for adults [\(ATSDR, 2003\)](#page-8-0). Diet provides most of the necessary selenium, preventing against the health effects of selenium deficiencies such as Keshan and Kashin–Beck diseases, which cause cardiovascular and joint problems, respectively. On the other hand, over-consumption of selenium, i.e., intake levels greater than $400 \,\mu g \,day^{-1}$, can lead to selenium toxicity, known as selenosis. Symptoms of selenosis include dermal effects such as loss of hair and nails as well as neurological effects such as numbness and paralysis ([ATSDR, 2003\)](#page-8-0). In addition to these issues, high levels of selenium have led to birth defects in birds ([Ohlendorf et al., 1986](#page-7-0)) and fish [\(Lemly, 1993](#page-7-0)). Although no selenium overdoses in humans have been directly linked to coal combustion emissions and the effects of selenium inputs into the ecosystem may not be immediately noticeable, selenium inputs from coal emissions need to be monitored since excess selenium can be difficult to remove ([Lemly,](#page-7-0) [1997](#page-7-0)). The varying species of selenium may have

different fates in the environment. For instance, selenium dioxide is water-soluble [\(Zhang, 2002](#page-8-0)), and can be dissolved in rainwater when released into the atmosphere. Also, species such as elemental Se and $SeO₂$ can condense at low temperatures and readily adsorb onto particles [\(Yan, 2004\)](#page-8-0). Through their deposition onto soil or vegetation, they have a direct path into the food chain when such soil or vegetation is consumed. Certain plants, such as Astragalus bisulcatus, are known to accumulate selenium in their leaves, causing toxicities in grazing animals [\(Pickering et al., 2000\)](#page-7-0). Knowledge of selenium's speciation will allow scientists to monitor selenium by predicting its final form, where it is deposited, and, ultimately, its effects on the environment.

Previous experimental and theoretical work on the arsenic and selenium species that may be released from coal combustion was used to determine the possible species that may exist as reactants in the atmosphere. By combusting coal, Senior and Bool concluded that both arsenic and selenium are vaporized in the combustion process. The vapors then condense onto fly ash particles during cooling or remain in gaseous form [\(Senior et al., 2000](#page-7-0)). [Yan](#page-8-0) [et al. \(1999\)](#page-8-0) completed equilibrium modeling through minimizing the Gibbs free energy of the system likely to determine the composition of arsenic, selenium, and mercury species released from coal combustion. The species, AsO, AsCl₃, AsH₃, As₂, As₄, H₂Se, SeO, SeO₂, SeCl₂, and elemental Se were all found to be possible stable species in the flue gases of coal combustion. Yan et al. later performed combustion simulations confirming the existence of Se, SeO, and $SeO₂$ in combustion flue gases [\(Yan et al., 2001, 2004](#page-8-0)). Thermodynamic modeling was also carried out by [Miller et al. \(2003\)](#page-7-0) to determine the speciation of trace elements in coal combustion where As_4O_{10} , As_4O_7 , As_4O_6 , As_4O_8 , AsO, and AsO_2 were found to be probable emission species. From these predicted species, As, AsO, Se, and SeO were chosen for atmospheric oxidation reaction calculations via OH and $HO₂$ radical species. Although these investigations provide insight into the possible arsenic and selenium species that are released from coal combustion processes, they provide no insight into the atmospheric lifetime of these pollutants once released from the stacks. The current investigation involves an examination of the atmospheric cycle of these species generated from coal combustion flue gases.

Theoretical calculations were carried out to enhance the knowledge of gaseous arsenic and selenium speciation within the atmosphere. Beginning with an investigation of the theoretical thermochemical data for several key reactions allows for increased understanding of the pathway of anthropogenic arsenic and selenium cycling through the biosphere. According to [Molina et al.](#page-7-0) [\(1996\)](#page-7-0), the free radicals OH and $HO₂$ play a major role in determining the lifetimes of various compounds in the atmosphere. For this reason, this work focuses on the oxidation of As and Se species by OH and $HO₂$ to then determine the relative stabilities of the oxidized forms of these species. Once these oxidized forms are recognized it can be determined how they might cycle through the biosphere, i.e., through rainwater, particulates, soil, freshwater, etc. For instance, understanding the oxidation pathways of As and Se species to the potentially water-soluble forms, As, As(III)O, As(III) O_2 , and Se O_2 , will be crucial in the development of effective control technologies.

2. Computational methodology

Calculations were carried out using the Gaussian 03 suite of programs ([Frisch et al., 2004](#page-7-0)). Basis sets incorporating relativistic effects for the inner electrons were explored through the use of small core relativistic effective core potentials (RECP) for arsenic and selenium. This basis set employs the relativistic ECP28MWB pseudopotential of the Stuttgart group ([Martin and Sundermann, 2001\)](#page-7-0) for both arsenic and selenium, with the respective energy-optimized (4s2p)/[3s2p] and (4s5p)/[2s3p] Gaussian-type orbital (GTO) valence bases optimized using multiconfiguration Dirac–Fock (MCDF) calculations. The second basis set employs a relativistic compact effective potential, RCEP28VDZ of [Stevens and Krauss \(1992\),](#page-7-0) which replaces 28 of arsenic's and 28 of selenium's atomic core electrons, derived from numerical Dirac–Fock wavefunctions using an optimizing process based upon the energy-overlap functional. Energy-optimized $(5s5p)/[2s2p]$ Gaussian type double- ς quality sp and triple- ς quality d functions were used, with the triple- ς d functions essential for describing the orbital shape changes that exist with d occupancy. Although both of these basis sets were employed for As and Se species in the basis set justification, through a balance of accuracy and computational expense, just the ECP28MWB pseudopotential of

Stuttgart group was used for As and a complete Pople basis set for Se. More specifically, a comparison of the theoretical predictions from the pseudopotentials to those of a complete basis set was carried out with the complete basis set calculations performed using the 6-311G* basis set. This extended triple-zeta Pople basis set, which includes an additional polarization function, was used for selenium, hydrogen, and oxygen.

3. Basis set justification

To determine the most accurate level of theory, six arsenic and selenium species were evaluated using four methods: CCSD, CCSD(T), QCISD, and QCISD(T) combined with each of the Pople basis sets: 6-311G, 6-311+G, 6-311G*, and 6-311+G*, as well as the Stevens ECP and ECP28MWB pseudopotential of the Stuttgart group. The resulting vibrational frequencies were compared to experimental values in order to find the level of theory with the lowest absolute error. The accurate prediction of the vibrational frequencies of the species in a system also closely predicts the energy of the system, which is crucial for the calculation of thermodynamic and kinetic properties. The QCISD level of theory along with the 6-311G* basis set was chosen for all species except those containing arsenic, for which the ECP28MWB pseudopotential of the Stuttgart group was selected.

3.1. Vibrational frequency comparison

[Table 1](#page-3-0) presents computed vibrational frequencies along with experimental values. From the comparison of the vibrations, the best level of theory was chosen. When compared with experiment, QCISD/6-311G* performed well (within 20 wavenumbers) for all selenium species. Most of the arsenic species were too computationally intensive to run without ECPs. When compared with experimental values, QCISD with the ECP28MWB pseudopotential of the Stuttgart group performed well for all arsenic species. The selenium species were also less computationally intensive with the ECPs, but reduced computation time did not compensate for the greatly reduced accuracy. At the chosen level of theory, comparison between predicted theoretical vibrational frequencies and available experimental literature data shows an average absolute error of 14 wave numbers.

^aCRC: Handbook of [Chemistry](#page-7-0) and Physics (1989).

 b [Herzberg](#page-7-0) (1951).</sup>

^c[Thermodynamics](#page-8-0) Source Database (2005).

Table 2Calculated and experimental bond lengths (in \AA) and bond angles (in deg)

Species Multiplicity Bond length (A) Experimental (cm^{-1}) Level of theory

[aEvenson](#page-7-0) et al. (1970) and [Mizushima](#page-7-0) (1972).

^bCRC: Handbook of [Chemistry](#page-7-0) and Physics (1989).

3.2. Equilibrium geometry comparison

[Table 2](#page-4-0) presents the computed geometries along with available experimental values. Calculated arsenic to oxygen bond lengths among all levels considered in this study overpredict the experimental distance by 0.0424 Å on average. At the chosen level of theory for arsenic, specifically for the AsO species, the absolute error from the experimental value is 0.0346 Å . A similar analysis for the

SeO species reveals an absolute error of 0.016 Å , while the average selenium to oxygen bond length, over all levels investigated, is over predicted by 0.0634 Å . All levels of theory suggest a bent arrangement for the $AsO₂$ and $SeO₂$, which is consistent with experimental findings. Correlation between the computed values at the chosen levels of theory against experimental data confirmed the chosen levels of theory worked well for these species.

Table 3

Reaction enthalpies, entropies, and equilibrium constants of arsenic and selenium reactions

Reaction	Parameter Units		Temperature (K)					
			223.15	253.15	298.15	400	500	600
(a) Arsenic reactions $As + OH \rightarrow AsOH$	ΔH ΔS K_{eq}	kcal mol ⁻¹ -29.71 cal mol K^{-1} -24.47	5.65(23)	-29.81 -24.87 1.99(20)	-29.94 -25.35 2.54(16)	-30.18 -26.06 6.25(10)	-30.36 -26.46 3.08(7)	-30.49 -26.71 1.87(5)
$As + HO_2 \rightarrow AsO + OH$	ΔH ΔS K_{eq}	kcal mol ⁻¹ -22.94 cal mol K^{-1}	1.02 4.90(22)	-22.90 1.17 1.07(20)	-22.84 1.38 1.12(17)	-22.72 1.74 6.23(12)	-22.62 1.96 2.08(10)	-22.55 2.09 4.69(8)
$As + OH \rightarrow AsO + H$	ΔH ΔS K_{eq}	$kcal \, mol^{-1}$ cal mol K^{-1}	7.99 -1.81 6.05(9)	8.00 -1.77 $5.13(-8)$	8.02 -1.69 $5.65(-7)$	8.10 -1.46 $1.80(-5)$	8.21 -1.21 $1.40(-4)$	8.35 -0.96 $5.61(-4)$
$AsO + HO_2 \rightarrow AsO_2 + OH \Delta H$	ΔS K_{eq}	$kcal \, mol^{-1}$ cal mol K^{-1} $\frac{1}{2}$ and $\frac{1}{2}$ and $\frac{1}{2}$	43.58 0.01	43.68 0.38	43.82 0.90 2.06 (-43) 2.37 (-38) 1.19 (-32) 1.91 (-24) 1.32 (-19)	44.14 1.83	44.42 2.46	44.65 2.89 $2.31(-16)$
$AsO + OH \rightarrow AsO2 + H$	ΔH ΔS K_{eq}	$kcal \, mol^{-1}$ cal mol K^{-1}	74.51 -2.84	74.57 -2.56	74.68 -2.16 $2.52 (-74)$ 1.14 (-65) 6.02 (-56) 5.52 (-42)	74.96 -1.36	75.25 -0.71 $8.89(-34)$	75.55 -0.17 $2.77(-28)$
(b) Selenium reactions $Se + OH \rightarrow SeOH$	ΔH ΔS K_{eq}	kcal mol ⁻¹ -51.14 cal mol K^{-1} -22.46	1.53(45)	-51.24 -22.87 1.76(39)	-51.38 -23.36 3.61(32)	-51.63 -24.08 8.83(22)	-51.81 -24.50 1.97(17)	-51.96 -24.76 3.27(13)
$Se + HO_2 \rightarrow SeO + OH$	ΔH ΔS K_{eq}	kcal mol ⁻¹ -25.07 cal mol K^{-1}	2.37 1.18(25)	-25.04 2.51 1.46(22)	-24.98 2.72 4.02(18)	-24.87 3.03 1.79(14)	-24.78 3.24 3.46(11)	-24.71 3.36 5.45(9)
$SeO + HO_2 \rightarrow SeO_2 + OH$ ΔH	ΔS K_{eq}	$kcal \, mol^{-1}$ cal mol K^{-1}	19.52 -5.78 7.19(17)	-19.48 -5.61 3.92(15)	-19.40 -5.33 6.68(12)	-19.25 -4.88 2.82(9)	-19.09 -4.53 2.26(7)	-18.95 -4.28 9.31(5)
$Se + OH \rightarrow SeO + H$	ΔH ΔS K_{eq}	$kcal \, mol^{-1}$ cal mol K^{-1}	5.83 -0.46 $1.53(-6)$	5.84 -0.43 $7.31(-6)$	5.86 -0.36 $4.97(-5)$	5.93 -0.16 $5.30(-4)$	6.03 $0.07\,$ $2.39(-3)$	6.16 0.30 $6.63(-3)$
$SeO + OH \rightarrow SeO2 + H$	ΔH ΔS $K_{\rm eq}$	$kcal \, mol^{-1}$ cal mol K^{-1} $\frac{1}{2}$ and $\frac{1}{2}$ and $\frac{1}{2}$	11.38 -8.61	11.40 -8.55	11.44 -8.40 9.31 (-14) 1.96 (-12) 8.26 (-11) 8.36 (-9)	11.56 -8.07	11.72 -7.70 $1.56(-7)$	11.92 -7.34 $1.33(-6)$

Numbers in parentheses indicate powers of 10.

4. Results and discussion

[Table 3a and b](#page-5-0) present the reaction enthalpies, entropies, and equilibrium constants for a selection of gas phase reactions involving arsenic and selenium. Five As and Se oxidation reactions were found to be spontaneous, having negative Gibb's free energy changes:

$$
As + OH \to AsOH,
$$
 (1)

$$
As + HO_2 \to AsO + OH,
$$
 (2)

$$
Se + OH \rightarrow SeOH, \tag{3}
$$

$$
Se + HO_2 \to SeO + OH,
$$
 (4)

$$
SeO + HO2 \rightarrow SeO2 + OH.
$$
 (5)

Of these, all were more likely to go to completion at tropospheric (lower) temperatures than at stratospheric (higher) temperatures. Graphs showing variations in K_{eq} over a range of temperatures appear in Figs. 1 and 2. For elemental arsenic, its oxidation pathway is more likely to be the formation of AsOH by OH in the troposphere and to be the formation of AsO by $HO₂$ in the stratosphere. For elemental selenium, its oxidation pathway is more likely to be oxidation by the OH radical forming SeOH than oxidation by the $HO₂$ radical forming SeO at all temperatures. For AsO, no oxidation pathways were thermodynamically favored. For SeO, oxidation by $HO₂$ was favored, whereas oxidation by OH was not thermodynamically favored. The reactions involving capture of the oxygen atom from the OH radical by As or Se were

Fig. 1. Equilibrium constant plot for arsenic reactions.

Fig. 2. Equilibrium constant plot for selenium reactions.

not thermodynamically favored. Equilibrium constants were evaluated over stratospheric pressures ranging from 0.2 to 1.0 atm; however, the order of magnitude of the equilibrium constants did not change with varying pressure. These thermodynamic calculations provide further understanding on the atmospheric speciation of arsenic and selenium species released from coal combustion processes. In future work a kinetic investigation will be carried out on the thermodynamically favored pathways to provide further insight into the atmospheric lifetimes of these coal combustion-generated pollutants.

5. Conclusions

Overall, the QCISD/6-311G* and QCISD with the ECP28MWB pseudopotential of the Stuttgart group performed best for all species and were used to calculate the energetics. From equilibrium constant calculations, possible oxidation–reduction mechanisms of As and Se compounds in the atmosphere were obtained. In particular, $SeO₂$ is water-soluble and the reaction between SeO and HO2 may present a way for selenium from coal combustion to end up in drinking water. More work on arsenic and selenium remains. Kinetic studies need to be carried out on the thermodynamically favored reactions in order to find which reactions actually take place under atmospheric conditions. Once the thermodynamically and kinetically favored species of arsenic and selenium are determined, their properties, including water solubility and adsorption potential, must be fully understood so that the life cycle of these species can be traced through the biosphere.

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